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Feedback effects between plant and flower-visiting insect communities along a primary succession gradient

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Abstract: Primary successions of glacier forelands are unique model systems to investigate community dynamics and assembly processes. However, successional changes of plant and insect communities have been mainly analysed separately. Therefore, changes in plant–insect interactions along successional gradients on glacier forelands remain unknown, despite their relevance to ecosystem functioning. This study assessed how successional changes of the vegetation influenced the composition of the flower-visiting insect assemblages of two plant species, *Leucanthemopsis alpina* (L.) Heyw. and *Saxifraga bryoides* L., selected as the only two insect-pollinated species occurring along the whole succession. In addition, we investigated the links between reproductive output of these plants and pollinator abundance through experimental exclusion of pollinators. Plant community structure changed along the succession, affecting the distribution and the abundance of insects via idiosyncratic responses of different insect functional groups. *L. alpina* interacted with ubiquitously distributed pollinators, while *S. bryoides* pollinators were positively associated with insect-pollinated plant species density and *S. bryoides* abundance. With succession proceeding, insect assemblages became more functionally diverse, with the abundance of parasitoids, predators and opportunists positively related to an increase in plant cover and diversity. The reproductive output of both plant species varied among successional stages. Contrary to our expectation, the obligate insect-pollinated *L. alpina* showed a reproductive output rather independent from pollinator abundance, while the reproductive output of the self-fertile *S. bryoides* seemed linked to pollinator abundance. Observing ecological interactions and using functional traits, we provided a mechanistic understanding of community assembly processes along a successional gradient. Plant community diversity and cover likely influenced insect community assembly through bottom-up effects. In turn, pollinators regulate plant reproductive output through top-down control. We emphasise that dynamics of alpine plant and insect communities may be structured by biotic interactions and feedback processes, rather than only be influenced by harsh abiotic conditions and stochastic events.

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**1 Feedback effects between plant and flower-visiting insect communities along a primary
2 succession gradient**

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25 **Abstract**

26 Primary successions of glacier forelands are unique model systems to investigate community
27 dynamics and assembly processes. However, successional changes of plant and insect
28 communities have been mainly analysed separately. Therefore, changes in plant–insect
29 interactions along successional gradients on glacier forelands remain unknown, despite their
30 relevance for ecosystem functioning. This study assessed how successional changes of the
31 vegetation influenced the composition of the flower-visiting insect assemblages of two plant
32 species, *Leucanthemopsis alpina* (L.) Heyw. and *Saxifraga bryoides* L., selected as the only
33 two insect-pollinated species occurring along the whole succession. In addition, we
34 investigated the links between reproductive output of these plants and pollinator abundance
35 through experimental exclusion of pollinators. Plant community structure changed along the
36 succession, affecting the distribution and the abundance of insects via idiosyncratic responses
37 of different insect functional groups. *L. alpina* interacted with ubiquitously distributed
38 pollinators, while *S. bryoides* pollinators were positively associated with insect-pollinated
39 plant species density and *S. bryoides* abundance. With succession proceeding, insect
40 assemblages became more functionally diverse, with the abundance of parasitoids, predators
41 and opportunists positively related to an increase in plant cover and diversity. The
42 reproductive output of both plant species varied among successional stages. Contrary to our
43 expectation, the obligate insect-pollinated *L. alpina* showed a reproductive output rather
44 independent from pollinator abundance, while the reproductive output of the self-fertile *S.*
45 *bryoides* seemed linked to pollinator abundance. Observing ecological interactions and using
46 functional traits we provided a mechanistic understanding of community assembly processes
47 along a successional gradient. Plant community diversity and cover likely influenced insect
48 community assembly through bottom-up effects. In turn, pollinators regulate plant

49 reproductive output through top-down control. We emphasize that dynamics of alpine plant
50 and insect communities may be structured by biotic interactions and feedback processes,
51 rather than only be influenced by harsh abiotic conditions and stochastic events.

52

53 **Keywords:** community assembly, functional traits, glacier forelands, plant–insect
54 interactions, reproductive output

55

56 **Introduction**

57 Glacier forelands are areas in front of retreating glaciers (Matthews 1992). They are the
58 product of the deglaciation since the end of the Little Ice Age (1500-1850 AD). Glacier
59 forelands represent unique model systems for studying ecosystem development (Walker et al.
60 2010) and community assembly processes along gradients of primary successions (Foster and
61 Tilman 2000). Following the time-sequence approach, it is assumed that plots of different
62 time since deglaciation represent different community development stages along a
63 successional gradient (Walker et al. 2010).

64 Successional changes of plant and insect communities were historically described
65 separately. Studies of primary succession along glacier forelands mainly focused on plant
66 colonization patterns (Matthews 1992; Erschbamer et al. 2008), on mechanisms of plant
67 succession related to environmental changes (Caccianiga et al. 2001; Hodkinson et al. 2003),
68 on plant species reproductive features (Erschbamer et al. 2001; Raffl et al. 2007; Jones and
69 del Moral 2009), and on plant functional traits (Caccianiga et al. 2006; Erschbamer and
70 Mayer 2011). From a zoological point of view, studies mainly focused on diversity patterns of
71 epigeal arthropod succession (e.g. Kaufmann 2001; Gobbi et al. 2006; Vater and Matthews
72 2013) and on community-environment relationships (e.g. Hodkinson et al. 2001; Gobbi et al.

73 2007).

74 However, little attention is given to plant–insect interactions along glacier foreland
75 successions (Albrecht et al. 2010), despite their role in governing community dynamics
76 (Losapio et al. 2015). Few recent studies showed common patterns in plant and ground beetle
77 functional traits (Gobbi et al. 2010), revealed the structure of plant–pollinator networks
78 (Albrecht et al. 2010; Losapio et al. 2015) and examined food web properties (König et al.
79 2011; Raso et al. 2014).

80 Despite the pivotal role played by plant–insect interactions for structuring communities
81 (Scherber et al. 2010), whether vegetation changes may influence the assembly of flower-
82 visiting insect communities along primary succession remains unexplored. Furthermore,
83 knowledge about the links between pollinator abundance and plant reproductive output along
84 primary succession gradients is still missing.

85 In the present study we aimed to describe feedbacks effects between plant and insect
86 communities along the primary succession of a glacier foreland (Vedretta d'Amola debris-
87 covered glacier, Adamello-Presanella Alps, Italy). In particular, we investigated how
88 vegetation changes along the succession influenced the composition of the flower-visiting
89 insect assemblages of two plant species, *Leucanthemopsis alpina* (L.) Heyw. and *Saxifraga*
90 *bryoides* L. In addition, we assessed the breeding systems (i.e. cross- vs self-pollination) of
91 these species and the links between plant reproductive output (i.e. number of fruits and seeds)
92 and pollinator abundance. Along the successional gradient we analysed: (i) changes in plant
93 cover, species diversity, species evenness, and plant pollination strategy; (ii) the distribution
94 and the abundance of insect assemblages on *L. alpina* and *S. bryoides* and their relationship
95 with the surrounding plant community structure, and (iii) the links between *L. alpina* and *S.*
96 *bryoides* reproductive output and pollinator abundance. Overall, we hypothesised that

97 successional changes of the vegetation can influence the functional composition of the flower-
98 visiting insect assemblages. In turn, we expected a link between plant reproductive output and
99 pollinator abundance.

100

101 **Materials and methods**

102 **Study area**

103 The study was performed along the primary succession of Vedretta d'Amola debris-covered
104 glacier foreland (Adamello-Presanella Alps, Italy, 46°13'16"N, 10°40'41"E). This glacier
105 foreland is 1.23 km long, facing south-east and extending from 2425 to 2560 m a.s.l. (Online
106 Resource Fig. 4). The glacier tongue position during the three main advance phases was dated
107 back to the end of the Little Ice Age (c. 1850 A.D.), 1925 A.D. and 1994 A.D. The nearest
108 climatic station in Pradalago at 2084 m a.s.l., about 9 km north-east from the study site,
109 receives a mean annual rainfall of 1395 mm and mean annual temperature is 2.6 °C. Snow cover
110 usually lasts from October to July. The glacier foreland has a patchy vegetation (c. 10-70%
111 plant cover) characterised by *Geum reptans* L., *Oxyria digyna* (L.) Hill, and *Luzula alpino-*
112 *pilosa* (Chaix) Breistr. Outside the glacier foreland, the vegetation is dominated by *Carex*
113 *curvula* All. and *Salix herbacea* L. (Online Resource Tab. 1).

114

115 *Study species*

116 We studied plant–insect interactions between the flowering plants *Leucanthemopsis alpina*
117 (L.) Heyw. (Asteraceae) and *Saxifraga bryoides* L. (Saxifragaceae) and their flower-visiting
118 insects. These two focal plant species were selected as they were the only two insect-
119 pollinated species occurring along the whole successional gradient (Online Resource Tab.1).

120 *L. alpina* is a perennial hemicryptophyte, forming leaf rosettes with prostrate stems and

121 deeply incised leaves (Online Resource Fig. 7). Reproductive shoots usually bear one terminal
122 capitulum (hereafter, inflorescence) consisting of yellow disk florets surrounded by white
123 ligulate ray florets. The fruit is an achene devoid of pappus. The species inhabits scree slopes,
124 grasslands and snowbeds from the subalpine to the nival zone (Pignatti 1982).

125 *S. bryoides* is a perennial semi-evergreen cushion plant forming densely matted shoots
126 with rigid, recurved leaves (Online Resource Fig. 8). Reproductive shoots usually bear one
127 terminal hermaphroditic white-yellowish flower with nectaries. The fruit is a bi-locular
128 capsule containing hundreds of tiny seeds. The species prefers siliceous substrates and grows
129 mainly in open habitats on scree and solid rock, from the alpine zone up to the nival zone
130 (Ladinig and Wagner 2007).

131

132 **Study design**

133 Along the successional gradient, five sites corresponding to the following successional stages
134 were selected: successional stage 0 (the debris-covered glacier surface), successional stage I
135 (1-20 years), successional stage II (21-90 years), successional stage III (91-160 years),
136 successional stage IV (>160 years). Each sampling site included two 25 m² plots (5m x 5m),
137 one for *L. alpina* and one for *S. bryoides*. Plots were established in areas not subjected to
138 disturbance.

139

140 *Sampling protocol*

141 Vascular plant species and bryophytes occurring within each 25 m² plot were recorded. The
142 overall plant cover and the abundance of every plant species were visually estimated with a
143 resolution of 5% (Online Resource Tab.1). Nomenclature follows Aeschimann et al. (2004).

144 For each plot, flower-visiting insects were collected during the 2012 and 2013 flowering

145 seasons (July and August). Flower-visiting insects were sampled with an entomological
146 aspirator, according to the time-observation plot method (Gibson et al. 2011; Nielsen et al.
147 2011). Within each plot, three sampling units were selected for each plant species, consisting
148 of an *L. alpina* tuft or an *S. bryoides* cushion, respectively (size about 20 cm diameter). The
149 three sampling units were surveyed together during periods of 40 minutes. Sampling was
150 repeated at 11am, 1pm, 3pm randomly across the season, resulting in 9 samples collected for
151 each plot at each successional stage every year. Due to long-lasting snow during 2013, *S.*
152 *bryoides* did not flower at the successional stage 0, while at stages I, II and III we collected 6
153 samples each. A total of 90 (45 for *L. alpina* and for *S. bryoides*, each) and 72 (45 and 27 for
154 *L.alpina* and *S. bryoides*, respectively) samples were carried out during 2012 and 2013,
155 respectively (Online Resource Tab. 2).

156 During each sampling occasion, the following weather conditions were recorded: (i) air
157 temperature and wind speed, at 0.5m above ground with a digital therm-anemometer (NTC
158 Thermometer Anemometer, ETANG Electronics, Birmingham, UK); (ii) cloud cover, visually
159 estimated with a resolution of 10% (100% as clear sky), and (iii) presence of gusts of wind
160 and/or rainfall (weather conditions during sampling reported in Online Resource Tab. 2). To
161 not undermine insect activity, we sampled from a distance of 0.5m away from plants and we
162 wore dull clothes.

163 Flower-visiting insects were identified to species level, when possible, otherwise to
164 genus or family level. Insect taxa were categorised into six functional groups, according to
165 their trophic guilds obtained from the literature: detritivores, phytophagous, pollinators,
166 predators, parasitoids, and opportunists (Mellini 1997; Gregor et al. 2002; Oosterbroek 2006;
167 Jedlička et al. 2009; Gobbi and Latella 2011, Speight 2013). Insect nomenclature follows de
168 Jong (2013).

169 To determine plant breeding systems (i.e. cross- vs self-pollination) and plant
170 reproductive output (i.e. number of fruits or seeds), a breeding experiment was conducted for
171 paired samples along the succession during 2013. Within each plot, five individuals per plant
172 species were randomly selected. For each individual, before anthesis one inflorescence/flower
173 was bagged with a closed fine-meshed nylon bag (hereafter treatment), while another
174 inflorescence/flower was bagged with an open bag (hereafter control). At the end of anthesis,
175 a closed bag was fixed on the control flower. Bags ($n = 92$; Online Resource Tab. 3) were
176 harvested in early October. Bags were stored at room temperature for 72h in paper before
177 fruits and seeds were counted.

178

179 **Statistical analysis**

180 *Plant community structure*

181 In order to examine how successional changes of the vegetation can influence the composition
182 of flower-visiting insect assemblages on the focal plant species, we first examined changes in
183 plant community structures along the succession. Thus, the following variables supposed to
184 describe successional changes in biodiversity and biomass of vegetation (Walker et al. 2010)
185 were considered to describe the plant community structure: plant cover, Shannon diversity,
186 Shannon evenness, insect-pollinated plant density and insect-pollinated plant richness. Insect-
187 pollinated plant density (e.p.d.) represents the relative abundance of insect-pollinated plant
188 species, while insect-pollinated plant richness (e.p.r.). represents the relative richness of
189 insect-pollinated plant species. Specifically:

$$e.p.d. = \sum_{i=1}^P c_i \times \sigma_i$$
$$e.p.r. = \frac{\sum_{i=1}^P s_i \times \sigma_i}{P}$$

190 where for the plant species i in a community with P species, c is the relative abundance (i.e.
191 relative cover) of the species i , s is species presence (i.e. presence/absence), σ is the
192 pollination strategy, with $\sigma = 1$ for insect-pollinated species and $\sigma = 0$ for wind-pollinated
193 species. These metrics have values ranging from 0 to 1. When a plant community is entirely
194 composed by insect-pollinated plants, e.p.d. and e.p.r. are equal to 1. Vice versa, when a plant
195 community is entirely composed by wind-pollinated plants, e.p.d. and e.p.r. are equal to 0.

196 To analyse the multivariate relationships among ecological indices of plant community
197 structure, a Principal Component Analysis (PCA) was carried out. Then, ANOVA-GLMs
198 were used to examine differences among successional stages (response, categorical variable
199 with five classes) for each ecological index (explanatory, continuous variables).

200

201 *Flower-visiting insect assemblages*

202 To assess differences of total insect abundance among successional stages, GLMs
203 (Poisson errors) were used, with insect abundance as response and successional stage as
204 explanatory variable. Separated analyses were performed for *L. alpina* and *S. bryoides*. Then,
205 the flower-visiting insect assemblages was examined focusing both on taxonomical and
206 functional composition (i.e. the six insect functional groups mentioned above). To examine
207 taxonomic and functional diversity changes along the succession, the Shannon index (at
208 family level) was calculated by pooling data for each successional stage. Pooled data were
209 then plotted along the successional gradient.

210 To describe the variation in insect assemblages along the successional gradient as well
211 as the associations among insect abundances and plant community structure, partial Canonical
212 Correspondence Analyses (pCCA) (ter Braak 1986; Jongman et al. 1995) were carried out.
213 Separated analyses were performed for *L. alpina* and *S. bryoides* assemblages, considering as

214 response variable the abundance of insects, at both taxonomical and functional levels. Thus,
215 in total four pCCAs were performed pooling insect abundance for each sampling unit (i.e.
216 plant tuft or cushion) for the two sampling years. Taxa and functional roles with a total
217 abundance lower than 5% were excluded. Thus, for the pCCAs using the taxonomical
218 information of insects, Coleoptera, Staphylinidae, Diptera Anthomiidae, Muscidae, Syrphidae
219 and Hymenoptera Ichneumonidae families were examined. For the pCCAs using the
220 functional groups of insects, pollinators, predators, opportunists, phytophagous and
221 parasitoids were considered. The explanatory 'environmental' variables were plant cover,
222 Shannon index, Shannon evenness, e.p.d. and e.p.r. The cover and the number of *L. alpina*
223 inflorescences and *S. bryoides* flowers were included as covariates, as they may potentially
224 influence flower visitors activity (Potts et al. 2003; Hegland and Boeke 2006). For all
225 explanatory variables a preliminary assessment of correlation was carried out (using the
226 Pearson correlation coefficient) in order to minimize multicollinearity issues in the pCCAs.
227 When the correlation between variables exceeded 0.75, only one of these variables was
228 included in the analysis, the choice based on subject-matter considerations (Legendre and
229 Legendre 2012). Since discontinuities in the sample distribution of plant cover and number of
230 *L. alpina* inflorescences were revealed, before performing the pCCAs these variables were
231 categorized as follows: plant community cover in two classes (<50, >50) and number of *L.*
232 *alpina* inflorescences in three classes (<10, 20-25, >30). Weather conditions during sampling
233 (mean temperature, wind speed, cloud cover and presence of strong wind gusts) were
234 considered as potential confounding effects, and included in the pCCAs as conditioning
235 variables (Oksanen et al. 2014). Results were reported in terms of proportions of explained
236 inertia as well as plotting the pCCA diagram.

237

238 *Plant breeding systems*

239 To identify the breeding systems of *L. alpina* and *S. bryoides*, seed set was analysed
240 considering: (i) Reproductive potential: number of ovaries per *L. alpina* inflorescence and
241 number of ovules per *S. bryoides* flower, and (ii) Reproductive output: number of *L. alpina*
242 fruits and number of *S. bryoides* seeds divided by reproductive potential. To assess
243 differences in reproductive potential and reproductive output between treatment and control
244 bags, conditional logistic regression models were used (Thernau 2014). Such a model allowed
245 us to account for the 'pairing' between treatment and control bags within the same plant.
246 Results were reported in terms of conditional odds ratio.

247 Finally, to link the reproductive output of *L. alpina* and *S. bryoides* with pollinator
248 abundance, the mean plant reproductive output the pollinator abundance (for the 2013 season)
249 were plotted together within each successional stage.

250 All statistical analyses were performed in R version 3.0.2. (R Core Team 2014).

251

252 **Results**

253 **Plant community structure**

254 Plant community structure varied along the successional gradient (Fig. 1). Sixty percent of the
255 variation was represented by the first principal component. This axis run from stage 0-I
256 communities, which had low plant cover (c. 10%), low Shannon diversity (2-2.4), high e.p.d.
257 (0.7-0.9), and high evenness (0.9-1), to stage III-IV communities, which had high plant cover
258 (c. 80-90%), high Shannon diversity (2.5-3), low e.p.d. (0.4-0.6), and low evenness (0.8-0.9).

259 Plant cover, Shannon diversity, evenness, and e.p.d. significantly varied among
260 successional stages (ANOVA-GLM χ^2 -test: $p < 0.001$ for each variable), while e.p.r. did not
261 show any significant variation (Table 1).

262

263 **Flower-visiting insect assemblages**

264 Within the recorded plant–insect interactions, we collected 474 specimens on *L. alpina*, 97
265 and 377 during 2012 and 2013 (Online Resource Tab. 5), respectively, and 440 specimens on
266 *S. bryoides* 197 and 243 during 2012 and 2013 (Online Resource Tab. 6), respectively.
267 Diptera, Hymenoptera and Coleoptera were the most abundant orders, represented by 21, 13
268 and 6 families, respectively (Online Resource Tab. 7).

269 The insect assemblage on *L. alpina* was dominated by anthomyiid flies (Diptera:
270 Anthomyiidae) and house flies (Diptera: Muscidae) with 45% and 34% of all insects,
271 respectively. Rove beetles (Coleoptera: Staphylinidae) and hoverflies (Diptera: Syrphidae)
272 were less abundant, both with 6%. The insect assemblage on *S. bryoides* was characterised by
273 house flies with 23%, anthomyiid flies with 18%, rove beetles with 16% and ichneumonids
274 wasps (Hymenoptera: Ichneumonidae) with 14%. Considering insect functional roles,
275 pollinators represented 84% of all insects interacting with *L. alpina*, while predators and
276 opportunists represented 9% and 5% of the assemblage, respectively. On *S. bryoides* 59%
277 were pollinators, 15% parasitoids, 14% opportunists and 5% phytophagous.

278 The abundance of flower-visiting insects significantly changed among successional
279 stages (ANOVA-GLM χ^2 -test: $p < 0.001$) both for *L. alpina* (Fig. 2a) and *S. bryoides* (Fig. 2b).
280 Furthermore, for both plant species, both the insect taxonomical and functional diversity
281 increased along the successional gradient (Fig. 2 c, d).

282 The variation in insect assemblages along the successional gradient in association with
283 plant community structure is presented in Fig. 3. Concerning the occurrence and the
284 abundance of insects interacting with *L. alpina* (Fig. 3a), weather conditions during sampling,
285 above all air temperature and cloud cover, explained 43.7% of the variation, while plant

community structure explained 28.9%. The first axis of the pCCA correlated mainly with e.p.r. ($r = 0.74$) and e.p.d. indices ($r = 0.40$). Higher abundance of Staphylinidae and Syrphidae was negatively associated with the first axis (weighted average scores = -1.31 and -0.53, respectively), i.e. with low e.p.r. and low e.p.d. values. Anthomyiidae and Muscidae did not show any specific association with environmental variables, mainly due to their ubiquitous distribution along the succession (weighted average scores for axis 1 = 0.00 and 0.28, respectively).

Concerning the occurrence and the distribution of insects interacting with *S. bryoides* (Fig. 3b), weather conditions during sampling explained 21.5% of the variation, while plant community structure explained 59.0%. The first axis of the pCCA positively correlated with plant cover ($r = 0.77$) and negatively correlated with e.p.d. ($r = -0.89$) and *S. bryoides* cover ($r = -0.72$). Higher abundance of Ichneumonidae was positively associated with the first axis (weighted average scores = 1.30), i.e. with high cover and both low e.p.r. and *S. bryoides* abundance. Anthomyiidae and Staphylinidae showed an opposite trend (weighted average scores for axis 1 = -0.46 and -0.44, respectively). A differentiation between successional stage IV and all other stages emerged, being the former characterised by higher abundance of Ichneumonidae.

Regarding the insect functional role distribution on *L. alpina* (Fig. 3c), weather conditions during sampling explained 6.8% of variation, while plant community structure explained 40.6%. The first axis was positively correlated with the e.p.r. index ($r = 0.81$) and negatively correlated with a high abundance of *L. alpina* inflorescences ($r = -0.66$) and plant cover ($r = -0.61$). Predators and opportunists were negatively associated with the first axis (weighted average scores = -0.73 and -0.93, respectively), i.e. they were positively associated with high plant cover and high *L. alpina* inflorescence abundance. Pollinators showed a

310 central position close to the axis origin (weighted average scores for axis 1 = 0.12). No
311 evident clustering of sites was found in both analyses.

312 Concerning the insect functional role distribution on *S. bryoides* (Fig. 3d), weather
313 conditions during sampling and plant succession variables explained 10.6% and 60.8% of the
314 variation, respectively. The first axis was positively correlated with plant cover ($r = 0.83$) and
315 negatively correlated with both e.p.d. ($r = -0.90$) and *S. bryoides* cover ($r = -0.76$). Pollinators
316 were positively associated with low plant cover, high e.p.d. and high *S. bryoides* cover
317 (weighted average scores for axis 1 = -0.42). On the other hand, parasitoids and opportunists
318 were positively associated with the first axis (weighted average scores = 1.04 and 0.60,
319 respectively), i.e. with high plant cover, low e.p.d. and low *S. bryoides* cover. The ordination
320 of successional stages reflected the main topographical sequence of the study sites. Stages 0
321 and I were characterised by high relative abundance of pollinator compared to parasitoids and
322 opportunists. On the other hand, stage IV was characterised by high plant cover, high
323 abundance of parasitoids and opportunists, and low pollinator abundance.

324

325 **Plant breeding system**

326 In total, 48 *L. alpina* inflorescences were collected and 3208 and 3391 ovaries were counted
327 for treatment and control bags, respectively. In total, 44 *S. bryoides* flowers were collected
328 and 2105 and 6017 ovules were counted for treatment and control bags, respectively (Online
329 Resource Tab. 8). The reproductive potential of *L. alpina* did not significantly change
330 between treatment levels (conditional odds ratio: 0.973; $p = 0.101$). For *S. bryoides* flowers, a
331 significant influence of bag treatment on the reproductive potential was found (conditional
332 odds ratio: 0.988; $p < 0.001$).

333 For *L. alpina*, the treatment effect on reproductive output was strong ($p < 0.001$), as

334 bagged inflorescences never produced fruits while control inflorescences produced on average
335 0.30 (± 0.22 SD) fruits per ovary (Online Resources Fig. 5a). For *S. bryoides*, the reproductive
336 output did not significantly change bag treatment levels (conditional odds ratio: 0.178; $p =$
337 0.140). Indeed, *S. bryoides* produced on average 0.16 (± 0.30 SD) and 0.29 (± 0.32 SD) seeds
338 per ovary for treatment and control bags, respectively. (Fig. 5b).

339 The mean reproductive output of *L. alpina* decreased along the succession towards early
340 successional stages, while its pollinator abundance increased (Online Resource Fig. 6a). The
341 reproductive success of *S. bryoides* and its pollinator abundance showed a coupled trend
342 among the successional stages, with an increase from stage 0 to stage II, followed by a
343 decrease from stage II to stage IV (Online Resource Fig. 6b).

344

345 **Discussion**

346 **Plant community structure**

347 The prevailing pollination strategy of plant communities changed from insect-pollinated to
348 wind-pollinated with succession proceeding, along with an increase in plant cover and plant
349 diversity. Specifically, the debris-covered glacier and the early successional stages were
350 characterised by scattered pioneer vegetation, which showed high evenness and a prevalent
351 pollination strategy dependent on insects. Wind-pollinated species increased in late-
352 successional stages, probably due to the increasing dominance of graminoids, which was
353 linked to an increase in plant cover.

354 Several studies on primary succession along glacier forelands showed that plant cover
355 (Matthews 1992), plant diversity (Caccianiga et al. 2001), community composition
356 (Hodkinson et al. 2003), functional traits (Caccianiga et al. 2006; Erschbamer and Mayer,
357 2011), and facilitation for recruitment and establishment (Erschbamer et al. 2008) change

358 along successional gradients. In the present study, we showed that also plant pollination
359 strategies changed along the succession. As a consequence, the observed change from
360 prevailing insect-pollinated to wind-pollinated plant species may produce cascading effects on
361 the assembly of insect community.

362

363 **Flower-visiting insect assemblages**

364 Pollinating flies were the dominant groups of insects visiting *L. alpina* along the succession.
365 This result is in accordance with several studies carried out in alpine environments across the
366 Alps (Albrecht et al. 2010; Wirth et al. 2011), Scandinavia (Elberling and Olesen 1999;
367 Totland 2003; Lázaro et al. 2013) and Patagonia (Devoto et al. 2005). On the other hand, *S.*
368 *bryoides* showed a diversified insect assemblage with pollinators, parasitoids, predators, and
369 opportunists unevenly distributed among successional stages.

370 The abundance, but also the the taxonomic and the functional diversity, of flower-
371 visiting insects increased along the succession. This pattern was consistent with the trends
372 emerged in previous studies for pollinators (Albrecht et al. 2010) and epigeal arthropod
373 communities (Kaufmann 2001; Gobbi et al. 2006; Gobbi et al. 2010; Vater and Matthews
374 2013) along other alpine glacier forelands. Furthermore, the observed patterns in flower-
375 visiting insects resembled the changes in plant community diversity. Therefore, this result
376 may indicate the presence of common and related mechanisms in the assembly of both plant
377 and insect communities (Losapio et al. 2015), as well as the direct role played by plant
378 succession in structuring the flower-visiting insect community.

379 Indeed, the observed change in plant community structure was the main source of
380 variation for the distribution and abundance of flower-visiting insects along the succession.
381 The most influential ecological factors were related to floral resource abundance, such as the

382 number of *L. alpina* inflorescences and *S. bryoides* cover, and to plant community diversity
383 and biomass, such as total plant cover and the density of insect-pollinated plants. These
384 results highlight the relevance of biotic interactions, in contrast with the generally minor role
385 played by abiotic conditions, for structuring communities along a glacier foreland succession.

386 Furthermore, different flower-visiting insect groups were differently related to
387 successional changes of the vegetation. This pattern suggests idiosyncratic changes of plant–
388 insect interactions along the successional gradient. Pollinators of *S. bryoides* were associated
389 with the density of insect-pollinated plants and to *S. bryoides* cover. As a consequence, the
390 abundance of floral resources at both community and population levels affected the
391 distribution and the abundance of *S. bryoides* pollinators. Other studies linking the abundance
392 of floral resources to the activity of pollinators showed that insects are strongly influenced
393 by floral density (Hegland and Boeke 2006). However, pollinators of *L. alpina* were
394 ubiquitously distributed, with low species turnover along the succession and independently
395 from *L. alpina* abundance and plant community structure. A possible explanation is that *L.*
396 *alpina* pollinators were generalists as they might feed on different species, so they did not
397 respond to changes in *L. alpina* floral abundance (Hegland and Boeke 2006) nor to
398 successional change of the vegetation.

399 Parasitoids, predators and opportunists characterised late successional stages and were
400 mainly related to an increase in plant cover. This is probably due to positive bottom-up effects
401 of plant community biomass and diversity on the abundance (Gobbi et al. 2010) and diversity
402 (Scherber et al. 2010) of resources and prey, which, in turn, might increase the abundance and
403 diversity of predators, parasitoids and opportunists.

404 Overall, our results suggest that successional changes of the vegetation are direct and
405 indirect drivers of the insect community assembly. If plant diversity and biomass influence

406 higher trophic levels through bottom-up trophic cascades (Scherber et al. 2010), in turn
407 pollinators may influence plant reproductive output (Albrecht et al. 2012) and plant
408 interactions (Lázaro et al. 2014) through top-down controls.

409

410 **Plant breeding system**

411 Based on our bag experiments we can conclude that *L. alpina* is strictly cross-pollinated, i.e. it
412 relies on insects for reproduction, while *S. bryoides* is self-pollinated, i.e. it does not require
413 insects for reproduction. In line with arctic-alpine plant reproductive strategies outlined by
414 Molau (1993), *L. alpina* had a pollen-risker profile, with high outbreeding rates and
415 reproduction resources allocated to pollen production and dispersion. On the other hand, *S.*
416 *bryoides* is an early-flowering seed-risker with high autogamy rates, in spite of the
417 proterandry as a phenological mechanism to enhance heterogamy (Ladinig and Wagner
418 2007), and resources allocated to seed production and recruitment.

419 The reproductive output of both plant species varied along the successional gradient. As
420 *L. alpina* requires pollinators for reproduction, it is expected to interact mainly with
421 pollinators and to show a reproductive output linked to pollinators availability (Molau, 1993).
422 Vice versa, as *S. bryoides* was able to self-fertilization, it is expected to interact with other
423 insect groups then pollinators and to show a reproductive output that is not limited by
424 pollinator availability. Consistently, *L. alpina* mainly interacted with pollinators while *S.*
425 *bryoides* was also interacting with parasitoids, predators and opportunists. However, contrary
426 to our expectation, the reproductive output of *L. alpina* was not linked with pollinator
427 abundance. This suggests that the reproductive output of this obligate insect-pollinated
428 species may also be influenced by other factors then pollinator abundance, such as the
429 heterospecific flower density of neighbouring plant community (Lázaro et al. 2013) or

430 competitive and facilitative interactions among co-flowering species for pollinators (Hegland
431 and Totland 2012). Surprisingly, there was a link between reproductive output of *S. bryoides*
432 and pollinator abundance. This result suggests that, for this species, out-crossing took priority
433 over selfing, with pollinator availability as an important factor for *S. bryoides* reproductive
434 output.

435

436 **Conclusion**

437 Our study is one of the few attempts to analyse plant–insect community feedbacks along a
438 successional gradient. Our approach offers a link between ecological interactions and
439 community assembly, revealing the relationships among different levels of ecological
440 organization. The key finding of our study is that structural changes of the vegetation during
441 succession influenced the distribution of flower-visiting insects. Although flower-visiting
442 insects would be able to cover the 1.2 km distance along the glacier foreland, the composition
443 of insect assemblages changed with succession proceeding due to guild-specific responses of
444 different insect functional groups to changes in plant community properties. Indeed, the
445 flower-visiting insect assemblage was initially dominated by pollinators, whose distribution
446 was related to the density and richness of insect-pollinated plants. With succession proceeding
447 insect assemblages became more diversified, with parasitoids, predators and opportunists
448 associated to the increase in plant cover and diversity. Overall, these findings provide an
449 important step forward towards a mechanistic understanding of the dynamics underlying the
450 assembly of interacting communities of different trophic levels. Our results suggests that the
451 assembly of alpine plant and insect communities may mainly be driven by biotic interactions
452 with inherent feedback effects, rather than only by harsh abiotic environmental conditions and
453 stochastic events. Thus, we emphasise the importance of biotic interactions underlying

454 ecological processes in alpine ecosystems.

455

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462

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582 **Tab. 1** Generalized linear models were used to test differences of ecological properties of the
 583 plant communities among successional stages. The significance of the predictor variable was
 584 assessed by comparing a model with and without the predictor.

Ecological index	χ^2	df	p
Plant cover	54.077	4	< 0.001
Shannon diversity	20.913	4	< 0.001
Shannon evenness	22.240	4	< 0.001
e.p.d.	22.737	4	< 0.001
e.p.r.	5.370	4	0.251

585

Tab. 2 pCCA synthesis results considering both taxonomic and functional features (level) of insect assemblages and as response variables abundance data of insects (i.e. number of individuals) pooled per sampling unit. Response variables: Ant = Anthomyiidae; Mus = Muscidae; Syr = Syrphidae; Stap = Staphylinidae; Icn = Ichneumonidae; Pol = Pollinators; Pred = Predators; Opp = Opportunists; Phy = Phytophagous; Par = Parasitoids. Variance indicate the percentage explained by explanatory variables. Shown are the most relevant explanatory variables (explanatory): Cov = plant community cover; e.p.r. = insect-pollinated plant richness; e.p.d. = insect-pollinated plant density; Leu = number of *L. alpina* inflorescences; Sax= *S. bryoides* cover.

Leucanthemopsis alpina

Level	Response	Variance [%]	Explanatory
Taxonomic	Ant, Mus, Syr, Stap	28.9	e.p.r.+e.p.d.
Functional	Pol, Pred, Opp	40.6	Leu + Cov + e.p.r.

Saxifraga bryoides

Level	Response	Variance [%]	Explanatory
Taxonomic	Ant, Mus, Stap, Icn	59.0	e.p.d. + Cov
Functional	Pol, Pred, Opp, Phyt, Par	60.8	Sax + e.p.d. + Cov

598 **Fig. 1** Biplot diagram showing the two principal axes of the plant community PCA based on
599 the relative cover values of all plant species. Plots and successional stages are represented
600 with different symbols, variable vectors (cover, Shannon diversity, evenness, e.p.d. and e.p.r.
601 indices) with arrows and successional stages with diamonds, connected by a line which
602 demonstrates their ranking according to the topographical sequence of the study sites. PC =
603 Principal component. The coordinate of the arrows relative to the top and right axes indicate
604 the correlation between variables and the first and the second principal axis, respectively.

605 **Fig. 2** Ecological indices of the flower-visiting insect community along the successional
606 gradient. Number of insect visits (**a, c**), taxonomic (dashed lines with circles) and functional
607 (solid lines with triangles) Shannon diversity index (**b, d**) for *L. alpina* (**a-b**) and *S. bryoides*
608 (**c-d**). Data were pooled for each successional stage.

609 **Fig. 3** pCCA of the *L. alpina* insect family abundance **(a)**. Variance was decomposed in
 610 43.7% weather conditions, 28.9% explanatory variables, and 27.4% unexplained. For the first
 611 axis net explained variance was 34.0%. and for the second axis 11.7%. pCCA of *S. bryoides*
 612 insect family abundance **(b)**. Variance was decomposed in 21.5% weather conditions, 59.0%
 613 explanatory variables, and 19.5% unexplained. For the first axis net explained variance was
 614 60.2% and for the second axis 9.5%. pCCA of the *L. alpina* insect functional roles abundance
 615 **(c)**. Variance was decomposed in 6.8% weather conditions, 40.6% explanatory variables, and
 616 52.6% unexplained. The first axis explained 35.0% of the net variance and the second axis
 617 8.6%. pCCA of *S. bryoides* insect functional roles abundance **(b)**. Variance was decomposed
 618 in 10.6% weather conditions, 60.8% explanatory variables, and 28.6% unexplained. For the
 619 first axis net explained variance was 57.0% and for the second axis 9.9%. Left axes are scaled
 620 proportional to species eigenvalues, top and right axes are scaled proportional to the
 621 correlation between the axis and the explanatory variable. The coordinates on the pCCA axis
 622 2 of four 'outlying' points (3 in panel c; 1 in panel d) were reduced to 3.5 (-3.5 when
 623 negative) in order to improve the readability of the diagrams. The coordinates of the arrows
 624 relative to the top and right axes indicate the correlation of the environmental variable
 625 mentioned above with the first and the second canonical axes. 'Environmental' explanatory
 626 variables: shan = plant community Shannon index; cov= plant cover; leu = *L. alpina* cover;
 627 sax= *S. bryoides* cover; epd = insect-pollinated plant density; epr = insect-pollinated plant
 628 relative richness; flowers= *S. bryoides* flowers. Centroid of categorical environmental
 629 variables (panel a and c): inf<10 = less than 10 inflorescences of *L. alpina*; inf20-25 = 20-25
 630 inflorescences of *L. alpina*; inf>30 = more than 30 inflorescences of *L. alpina*; cov<50 = plant
 631 cover lower than 50%, cov>70 = plant cover higher than 50%.